



Policy-driven analysis of carbon emission efficiency under uncertainty and its application in Chinese industry: Hybrid delta-slacks-based model and ordinal priority approach

Shutian Cui^a, Renlong Wang^b , Xiaoyan Li^a, Xiuguang Bai^{a,*}

^a School of Economics and Management, Northwest A&F University, Yangling, 712100, China

^b School of Emergency Management Science and Engineering, University of Chinese Academy of Sciences, Beijing, 100049, China

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ABSTRACT

Regional differences in carbon emission efficiency arise from disparities in multiple factors, which are often influenced by government policy preferences. However, currently, most studies fail to consider the impact of government policy preferences and data uncertainty on carbon emission efficiency. To address the above limitations, this study proposes a hybrid model based on delta-slacks-based model and ordinal priority approach for measuring carbon emission efficiency driven by government policy preferences under data uncertainty. The proposed model incorporates constraints on the importance of inputs and outputs under different policy preference scenarios. It then develops the efficiency optimization model with Farrell frontiers and efficiency tapes to deal with the data uncertainty in inputs and outputs. This study demonstrates the proposed model by analyzing industrial carbon emission efficiency in Chinese provinces in 2021. It examines the carbon emission efficiency and corresponding clustering results of provinces under three types of policies with varying priority preferences. The results indicate that the carbon emission efficiency of the 30 provinces can mainly be categorized into technology-driven, development-balanced, and transition-potential types, with most provinces achieving optimal efficiency under the technology-dominant preferences. Ultimately, this study suggests policy recommendation for different provinces to work towards achieving the low carbon goal.

1. Introduction

The increasing emissions of greenhouse gases, represented by carbon dioxide, are exacerbating global climate change [1,2]. China, the world's largest emitter of carbon dioxide, actively engages in international climate cooperation, taking responsibility for global emission reduction amid significant pressure to cut emissions and conserve energy [3]. At the 75th session of the United Nations General Assembly, the Chinese government pledged to “strive to peak carbon dioxide emissions before 2030 and achieve carbon neutrality before 2060”, i.e., the dual carbon strategy [4]. Under the premise of sustained economic growth, how to effectively promote carbon emission efficiency (CEE) has become an essential issue that China needs to address urgently. Currently, China's industrial sector contributes 40.1% of GDP, but its energy consumption and carbon emissions account for 67.9% and 84.2% of the national total, respectively [5]. Improving industrial carbon emission efficiency (ICEE) has become crucial to achieving the dual carbon strategy. Notably, there are significant regional differences

in industrial sectors in terms of energy usage, operational efficiency, emission reduction potential, and technological levels, which leads to significant heterogeneity in efficiency across regions [6]. Therefore, the government must develop region-specific evaluation systems for ICEE based on the regional endowment, which usually reflects the policy preference. Specifically, policy preference refers to the government tendency to focus on particular objective or interests when prioritizing policies [7]. This tendency is essential in guiding government decisions on resource allocation, policy implementation, and monitoring. Policy preferences also reflect the degree of integration and importance governments attach to individual national development strategies in policy formulation. Studies have shown that the policy preferences of government can significantly impact carbon emissions [8]. Changes in these preferences, along with shifts in industrial structure and government interventions, profoundly affect the efficacy of reducing carbon emissions. Policy preferences directly shape the prioritization of policy

* Corresponding authors.

E-mail addresses: cui@nwfau.edu.cn (S. Cui), wangrenlong23@mails.ucas.ac.cn (R. Wang), lixiaoyan@nwfau.edu.cn (X. Li), baixg960@nwfau.edu.cn (X. Bai).

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implementation. The significance of inputs and outputs varies when assessing CEE under different policies.

1.1. Related literature

Studies related to CEE evaluation have mainly gone through a progression from single factor analysis to total factor analysis. Economically, the total output to total input factors ratio is typically employed to assess CEE. CEE was originally a single-factor measure defined by Yoichi Kaya et al. [9] as GDP divided by carbon emissions over time. Subsequent research has introduced many additional indicators, including the carbon index, carbon intensity, energy intensity, and emissions per capita per unit of GDP [10]. However, the single-factor approach is inadequate for capturing the multidimensional aspects of CEE. Consequently, a total factor approach has emerged, incorporating labor scale, capital inputs, and energy consumption to provide a comprehensive view of CEE [11,12]. Since the introduction of the total factor concept to energy efficiency measurement, it has gained prominence in academia. Methods for analyzing CEE are broadly categorized into parametric and non-parametric methods. However, parametric methods require a predetermined production function, which presents a practical challenge [13]. Thus, this study focuses on the non-parametric method, specifically DEA and its extensions, which currently dominates research in this field. Table A.1 outlines the critical literature on non-parametric methods for assessing CEE.

Table A.1 illustrates that recent studies have evaluated CEE with decision-making units (DMUs) across national, regional, industry, provincial, and municipal levels, considering regular, embedded carbon emissions, water pollution and carbon neutrality, and coordinated governance perspectives. The primary data envelopment analysis (DEA)-based non-parametric methods for CEE include radial, non-radial, and directional distance functions. In terms of the radial model, Liu et al. [14] applied the BCC model to assess changes in industrial eco-efficiency across 16 prefecture-level cities in Anhui, China from a static viewpoint. Ding et al. [15] performed a comparative analysis of CEE among 30 provinces in China using the CCR and BCC models. However, these conventional radial models overlook the selection of radial direction in efficiency measurement and encounter issues with slack efficiency measurement [16]. To address these issues, several studies have utilized the non-radial super-SBM model to measure CEE. For example, Jiang et al. [17] applied super-SBM to evaluate CEE in the logistics industry across 12 pilot regions in China. Gao et al. [18] integrated the trade openness factor into the embedded carbon emission perspective and employed super-SBM to analyze CEE across 28 industrial sectors in China. Jiang et al. [19] measured the CEE of 30 cities in Northwest China from 2011 to 2020 using a super-efficient SBM model based on the dual perspectives of water pollution and carbon neutrality. Fang et al. [20] utilized super-SBM to assess CEE at 42 thermal power plants in China in 2020 from a microscopic perspective. Meanwhile, to enhance environmental efficiency assessment incorporating undesirable outputs, Chung et al. [21] introduced the radial DDF based on Shepherd's approach. However, the radial DDF fails to eliminate inefficiencies caused by input and output slack, potentially leading to overestimating CEE. Färe and Grosskopf [22] introduced a generalized NDDF for total factor energy productivity, relaxing the requirement for desired and undesirable outputs to vary proportionally. Fukuyama and Weber [23] developed the SBM-DDF model for CEE, which integrates undesirable outputs to mitigate radial and directional biases. Moreover, some studies have proposed a multi-stage DEA model combining parametric and non-parametric approaches [24]. Among them, the most representative is three-stage DEA, which is capable of incorporating environmental factors and random noise in the assessment of efficiency [25].

However, it is noteworthy that the assessment of CEE depends heavily on the value judgments that policymakers make about the resource allocation scenarios and the future of the economy and the

environment [26]. This process highlights the potential impact of policy preferences on CEE, which refers to the specific preferences or prioritized objectives the government holds when formulating policies or selecting options. In the assessment of CEE, policy preferences can influence local behavior and decision-making through a variety of mechanisms that promote the transition of the industrial sector towards a higher level of efficiency and cleaner production patterns [27]. The empirical study conducted by Meng et al. [28] revealed significant discrepancies in the carbon emission performance of the manufacturing sector when subjected to scale-oriented and innovation-oriented carbon reduction policy preferences. Therefore, a profound comprehension and rigorous consideration of policy preferences is essential to assess alterations in CEE with precision. Such an analysis will assist the government in formulating more effective carbon emission reduction policies, considering the varying circumstances of different regions. Nevertheless, only a limited number of studies that assess CEE take policy preferences into account. In addition, the potential implications of data uncertainties should be considered, which may arise from factors such as statistical inaccuracies or human interference. These factors could significantly influence the assessment of CEE based on policy preferences, which represents a limitation of the current research [29].

To overcome the above limitations, the objective of this study is to propose a model for measuring CEE from the total factor perspective that can accommodate the various policy preference scenarios and account for potential data uncertainty.

1.2. Contributions

This study proposes a hybrid model based on delta-slacks-based model (δ -SBM) and ordinal priority approach (OPA) for CEE analysis under policy preferences and data uncertainty. Using the proposed model, this study examines the efficiency differences under economic, environmental, and technological priority policies and their preference scenario through an illustrative demonstration of the ICEE across 30 provinces of China in 2021. Then, K-means clustering is employed to analyze the weight frontiers of inputs and outputs across provinces under various policy scenarios, identifying groups with similar characteristics. Benchmark provinces in each category are identified by comparing the optimal efficiencies of provinces in each group. Finally, this study offers policy recommendations for different provinces to achieve carbon emission reduction, offering practical guidance for industrial sectors.

The primary contribution of this study lies in proposing hybrid δ -SBM-OPA model for analyzing CEE that can address policy preferences and data uncertainties within the decision-making scenario. Specifically:

- Methodologically, the proposed model formulates policy preferences as scenario constraints, thereby determining the efficiency of DMUs under the specific policy preference scenarios based on human judgment and actual input and output data. Unlike extended DEA models that rely on predetermined weights, the proposed model can be regarded as an integrated approach that narrows the feasible region of the original model through policy preferences, subsequently optimizing a set of optimal weights. Furthermore, the proposed model establishes efficiency tapes for inputs and outputs facing data uncertainty and calculate corresponding sensitivity indicators to better distinguish the efficiency of individual DMUs, providing more reliable analytical results.
- Practically, the proposed model offers policymakers a customizable and robust tool for analyzing CEE. Policymakers can observe changes in the CEE of DMUs under different policy scenarios by setting various policies with specific preferences. Conducting cluster analysis on optimal efficiencies and the weight frontier of inputs and outputs under different scenarios can reveal the optimal scenarios and developmental paths for DMUs to formulate relevant policies.

1.3. Organization

The remainder of this paper is organized as follows: Section 2 introduces the preliminary related to the basic model, i.e., OPA and δ -SBM. Section 3 proposes the δ -SBM-OPA model. Section 4 demonstrates the proposed model with a case study analyzing ICEE among Chinese provinces in 2021, considering specific policy preferences. Section 5 further discusses the variations in ICEE among Chinese provinces under different policy preferences and offers corresponding policy recommendations. Finally, Section 6 summarizes the findings and future directions.

2. Preliminary

This section outlines the preliminaries of OPA and δ -SBM, forming the foundation for the proposed hybrid δ -SBM-OPA model.

2.1. Ordinal priority approach

OPA is a promising multi-criteria decision-making (MCDM) technique under incomplete information [30]. The method applies across diverse contexts of MCDM, encompassing the determination of weights for experts, criteria, and alternatives in group and individual decision-making [31]. The strength of OPA lies in its utilization of more stable and readily accessible ranking data as inputs, thereby obtaining the weights for experts, criteria, and alternatives simultaneously through solving a linear programming model [32,33]. OPA has found extensive application in domains such as supplier selection [34], portfolio selection [35], performance evaluation [36], and project planning [37,38]. In this study, OPA will be utilized to incorporate the contextual constraints of policy preference for δ -SBM, restricting the impact of varying policy preferences on the efficiency of DMUs.

Consider a problem where a decision-maker must assign weights to rank a set of criteria, $\mathcal{J} := \{1, 2, \dots, J\}$, indexed by j , based on the evaluation of a set of experts, $\mathcal{K} := \{1, 2, \dots, K\}$, indexed by k . The initial step of OPA involves determining the rank of expert $s_k \in [K]$ for all $k \in \mathcal{K}$, considering aspects like domain expertise, professional experience, job titles, and positions. Subsequently, each expert $k \in \mathcal{K}$ independently assigns rank to criteria $r_{jk} \in [J]$ for all $j \in \mathcal{J}$ based on their own judgment and preferences. To facilitate the following discussion, define the following three sets:

$$\mathcal{A}^1 := \{(j, k, l) \in \mathcal{J} \times \mathcal{K} \times \mathcal{K} : r_{jl} = r_{jk} + 1, r_{jk} = 1, 2, \dots, J-1\},$$

$$\mathcal{A}^2 := \{(j, k) \in \mathcal{J} \times \mathcal{K} : r_{jk} = J\},$$

$$\mathcal{A} := \{(j, k) \in \mathcal{J} \times \mathcal{K}\}.$$

Intuitively, \mathcal{A}^1 contains the indices of criteria with consecutive rankings for each expert, while \mathcal{A}^2 contains the indices of criteria ranked given by each expert, and $\mathcal{A} := \mathcal{A}^1 \cup \mathcal{A}^2$ represents the indices of all experts and criteria. OPA seeks to maximize weight disparities that reflect expert preferences within the normalized weight space, through the following optimization model:

$$\begin{aligned} & \max_{w, Z} Z, \\ & \text{s.t. } s_k r_{jk} (w_{jk} - w_{jl}) \geq Z, \quad \forall (j, k, l) \in \mathcal{A}^1, \\ & \quad s_k r_{jk} (w_{jk}) \geq Z, \quad \forall (j, k) \in \mathcal{A}^2, \\ & \quad \sum_{k \in \mathcal{K}} \sum_{j \in \mathcal{J}} w_{jk} = 1, \\ & \quad w_{jk} \geq 0, \quad \forall (j, k) \in \mathcal{A}, \end{aligned} \quad (1)$$

where Z represents the weight disparity scalar and w_{jk} denotes the weight of j based on the evaluation of expert k . After solving Eq. (1), the weights of criteria w_j for all $j \in \mathcal{J}$ are calculated according to Eq. (2).

$$w_j = \sum_{k \in \mathcal{K}} w_{jk} \quad \forall j \in \mathcal{J} \quad (2)$$

2.2. Delta-slacks-based model

DEA, a non-parametric data analysis method, primarily assesses the performance of DMUs with multiple inputs and outputs [39]. Traditional DEA models, such as the CCR, BCC, ADD, and SBM, are significantly affected by the number of DMUs and the inputs and outputs [40]. As the number of DMUs decreases or inputs and outputs increase, the discriminative ability of traditional DEA models in evaluating DMU efficiency diminishes, tending to allocate more DMUs to technical efficiency scores. In practice, many input and output data exhibit a certain degree of uncertainty. This uncertainty primarily arises from factors such as information loss, knowledge constraints, and human errors, particularly in the field of carbon emissions data statistics [7,41]. The traditional DEA model fails to address efficiency evaluations when dealing with imprecise input and output data [42]. To overcome these limitations, Khezrimotlagh et al. [43] introduced the δ -SBM model. It is a type of robust DEA model, exhibiting higher flexibility by rational adjustments to the Farrell frontier of inputs and outputs. It creates an effective band region that distinguishes the efficiency levels among various DMUs. Therefore, this study mainly focuses on δ -SBM as the main body of the proposed model for evaluating the efficiency of the DMUs with imprecise input and output data under policy preference.

Suppose there is a set of DMUs, $\mathcal{I} := \{1, 2, \dots, I\}$, indexed by i , each with input vector $x_i \in \mathbb{R}^{J^-}$ and output vector $y_i \in \mathbb{R}^{J^+}$, where the sets of input and output criteria are $\mathcal{J}^- := \{1, 2, \dots, J^-\}$ and $\mathcal{J}^+ := \{1, 2, \dots, J^+\}$. Let w_j^- and w_j^+ denote the given weights assigned to input $j \in \mathcal{J}^-$ and output $j \in \mathcal{J}^+$, respectively. Given a degree of freedom ε for adjusting the Farrell frontier by shifting the inputs and outputs up or down, the δ -SBM model for evaluating the performance of each DMU $l \in \mathcal{I}$ is formulated as:

$$\begin{aligned} & \max_{\lambda, s^-, s^+} \sum_{j \in \mathcal{J}^-} w_j^- s_{lj}^- + \sum_{j \in \mathcal{J}^+} w_j^+ s_{lj}^+, \\ & \text{s.t. } \sum_{i \in \mathcal{I}} \lambda_i x_{ij} + s_{lj}^- = x_{lj} + \varepsilon_j^-, \quad \forall j \in \mathcal{J}^-, \\ & \quad \sum_{i \in \mathcal{I}} \lambda_i y_{ij} - s_{lj}^+ = y_{lj} + \varepsilon_j^+, \quad \forall j \in \mathcal{J}^+, \\ & \quad x_{lj} - s_{lj}^- \geq 0, \quad \forall j \in \mathcal{J}^-, \\ & \quad y_{lj} + s_{lj}^+ - 2\varepsilon_j^+ \geq 0, \quad \forall j \in \mathcal{J}^+, \\ & \quad \lambda_i, s_{lj}^-, s_{lj}^+ \geq 0, \quad \forall i \in \mathcal{I}, j \in \mathcal{J}^- \cup \mathcal{J}^+, \end{aligned} \quad (3)$$

where s_{lj}^- and s_{lj}^+ are the slack variables for input $j \in \mathcal{J}^-$ and output $j \in \mathcal{J}^+$ of DMU l , λ_i is the multiplier used to compute the linear combination of inputs and outputs for DMU i , and $\varepsilon_j^- = \varepsilon x_{lj}$ and $\varepsilon_j^+ = \varepsilon y_{lj}$ are allowed errors with ε -degree of freedom for inputs $j \in \mathcal{J}^-$ and outputs $j \in \mathcal{J}^+$ of DMU l .

After solving Eq. (3) for the optimal solution $(\lambda^*, s^{*-}, s^{*+})$, the best technical efficient target and score of DMU l with ε -degree of freedom can be represented as Eqs. (4) and (5).

$$\begin{cases} x_{lj}^* = x_{lj} - s_{lj}^{*-} + \varepsilon_j^- \\ y_{lj}^* = y_{lj} + s_{lj}^{*+} - \varepsilon_j^+ \end{cases} \quad (4)$$

$$\gamma_l = \frac{\sum_{j \in \mathcal{J}^+} w_j^+ y_{lj}^* / \sum_{j \in \mathcal{J}^-} w_j^- x_{lj}^*}{\sum_{j \in \mathcal{J}^+} w_j^+ y_{lj}^* / \sum_{j \in \mathcal{J}^-} w_j^- x_{lj}^*} \quad (5)$$

The lower and upper bound of efficient target of DMU l with ε -degree of freedom can be represented as Eqs. (6) and (7), respectively.

$$\begin{cases} x_{lj}^+ = x_{lj} - s_{lj}^{*-} + 2\varepsilon_j^- \\ y_{lj}^+ = y_{lj} + s_{lj}^{*+} - 2\varepsilon_j^+ \end{cases} \quad (6)$$

$$\begin{cases} x_{lj}^- = x_{lj} - s_{lj}^{*-} \\ y_{lj}^- = y_{lj} + s_{lj}^{*+} \end{cases} \quad (7)$$

The sensitivity score of DMU i for the uncertainty efficiency tape with ε -degree of freedom is shown in Eq. (8).

$$\eta_i = \frac{\sum_{j \in J^+} w_j^+ y_{ij}^+ / \sum_{j \in J^-} w_j^- x_{ij}^+}{\sum_{j \in J^+} w_j^+ y_{ij}^- / \sum_{j \in J^-} w_j^- x_{ij}^-} \quad (8)$$

Notably, the assigned weight w_j^- of each input variable can be defined as $1/\min_{i \in I} \{x_{ij}\}$, $1/\max_{i \in I} \{x_{ij}\}$ or $1/\text{avg}_{i \in I} \{x_{ij}\}$ and so on. And the weights w_j^+ of each output variables can be defined in the same way of y_{ij} . However, when setting weights for inputs and outputs, it essentially involves non-dimensional standardization in the original δ -SBM model, overlooking the subjective judgment and preference of decision-makers [36,44]. This factor might lead to impractical solutions, particularly when considering impact of varying policy preferences on CEE analysis. Hence, it becomes necessary to incorporate weights w_j^- and w_j^+ as variables of the δ -SBM model, originating from the decision-makers' preference perspective. In the next section, the δ -SBM-OPA model will be proposed for carbon emission analysis under policy preference with the aid of Eqs. (1) and (3).

3. The proposed model for carbon efficiency analysis under policy preference

This study focuses on CEE analysis under policy preference, where policymakers provide their preferences for policies along with the corresponding inputs and outputs. Consider a set of DMUs, $I := \{1, 2, \dots, I\}$, indexed by i , that need to be evaluated for CEE based on their performance $x_i \in \mathbb{R}^{J^-}$ on a set of inputs $J^- := \{1, 2, \dots, J^-\}$ and $y_i \in \mathbb{R}^{J^+}$ on a set of outputs $J^+ := \{1, 2, \dots, J^+\}$. Let $\mathcal{K} := \{1, 2, \dots, K\}$ represent the set of government policies, indexed by k . Policymakers first rank the policies $s_k \in [K]$ for each policy $k \in \mathcal{K}$ based on their preferences. Subsequently, the rankings of criteria (i.e., inputs and outputs) $r_{jk} \in [J]$ for all criteria $j \in J^- \cup J^+$ under policy $k \in \mathcal{K}$ are provided based on policymakers evaluations or data analysis. In this context, the inputs and outputs of the DMUs serve as the basic data for traditional CEE, while the policy rankings and the rankings of inputs and outputs under different policies that reflects the policy preference are the additional data required for the proposed model.

In this section, a hybrid δ -SBM-OPA model is proposed to analyze the CEE under multiple scenarios with different government policy preferences. The first step is to derive the dual problem of the original δ -SBM model since it offers lucid guidance on the weightage information to criteria (i.e., inputs and outputs). These weights delineate how each DMU prioritizes its inputs and outputs (e.g., capital inputs, industrial output, and carbon emissions) when striving for optimal efficiency of carbon emission considering policy preference. Moreover, the transformation of a dual problem converts the initial nonlinear optimization problem into a linear optimization and reduces the number of decision variables involved. The dual problem of the original δ -SBM model is shown in Eq. (9).

$$\begin{aligned} \min_{v^-, u^+, \theta^-, \sigma^+} \quad & \sum_{j \in J^-} (x_{ij} + \varepsilon_j^-) v_j^- + \sum_{j \in J^-} x_{ij} \theta_j^- \\ & - \sum_{j \in J^+} (y_{ij} + \varepsilon_j^+) u_j^+ - \sum_{j \in J^+} (y_{ij} - 2\varepsilon_j^+) \sigma_j^+ \\ \text{s.t.} \quad & \sum_{j \in J^-} x_{ij} v_j^- - \sum_{j \in J^+} y_{ij} u_j^+ \geq 0 \quad \forall i \in I \\ & v_j^- + \theta_j^- \geq w_j^- \quad \forall j \in J^- \\ & u_j^+ + \sigma_j^+ \geq w_j^+ \quad \forall j \in J^+ \\ & \theta_j^- \geq 0, \sigma_j^+ \leq 0 \quad \forall j \in J^- \cup J^+ \end{aligned} \quad (9)$$

Define the following sets:

$$\begin{aligned} \mathcal{A}^1 &:= \{(j, k, l) \in (J^- \cup J^+) \times \mathcal{K} \times (J^- \cup J^+) : r_{lk} = r_{jk} + 1, \\ & r_{jk} = 1, 2, \dots, J^- + J^+ - 1\}, \\ \mathcal{A}^2 &:= \{(j, k) \in (J^- \cup J^+) \times \mathcal{K} : r_{jk} = J^- + J^+\}. \end{aligned}$$

Eqs. (1) and (9) are integrated into a multi-objective optimization model, illustrated in Eq. (10), to account for the influence of policy preferences on the importance of inputs and outputs in analyzing CEE.

$$\begin{aligned} \min_{w, Z} \quad & \sum_{j \in J^-} (x_{ij} + \varepsilon_j^-) v_j^- + \sum_{j \in J^-} x_{ij} \theta_j^- \\ & - \sum_{j \in J^+} (y_{ij} + \varepsilon_j^+) u_j^+ - \sum_{j \in J^+} (y_{ij} - 2\varepsilon_j^+) \sigma_j^+ \\ \max_{w, Z} \quad & Z \\ \text{s.t.} \quad & \sum_{j \in J^-} x_{ij} v_j^- - \sum_{j \in J^+} y_{ij} u_j^+ \geq 0 \quad \forall i \in I \\ & v_j^- + \theta_j^- \geq \sum_{k \in \mathcal{K}} w_{jk} / \max_{i \in I} \{x_{ij}\} \quad \forall j \in J^- \\ & u_j^+ + \sigma_j^+ \geq \sum_{k \in \mathcal{K}} w_{jk} / \min_{i \in I} \{y_{ij}\} \quad \forall j \in J^+ \\ & s_k r_{jk} (w_{jk} - w_{lk}) \geq Z \quad \forall (j, k, l) \in \mathcal{A}^1 \\ & s_k r_{jk} (w_{jk}) \geq Z \quad \forall (j, k) \in \mathcal{A}^2 \\ & \sum_{k \in \mathcal{K}} \sum_{j \in J^- \cup J^+} w_{jk} = 1 \\ & \theta_j^-, w_{jk}^+ \geq 0, \sigma_j^+ \leq 0 \quad \forall j \in J^- \cup J^+, k \in \mathcal{K} \end{aligned} \quad (10)$$

where $\varepsilon_j^+ = 1/\max_{i \in I} \{x_{ij}\}$ and $\varepsilon_j^- = 1/\min_{i \in I} \{x_{ij}\}$. In Eq. (10), the first objective function and the first constraint belong to the dual problem of δ -SBM in Eq. (9). The second objective function and the fourth, fifth, and sixth constraints belong to OPA in Eq. (1). The right-hand side of the second and third constraint is the output of OPA in Eq. (1), and the left-hand side is the input of δ -SBM in Eq. (9). From a modeling perspective, OPA provides δ -SBM with lower bound constraints on the importance of inputs and outputs that take policy preferences into account.

Eq. (10) presents a multi-objective optimization model, which can be solved through various methods like Pareto-optimality, goal programming, budgeted-constraint approach, and the max-min approach [36,45,46]. This study utilizes the weighted max-min approach due to its flexibility for policymakers to express the relative importance between δ -SBM and OPA. Since the scales differ between δ -SBM and OPA, Eq. (11) is utilized to transform the objective functions into non-dimensional counterparts, with values fall within the range of [0,1].

$$f_k^{\text{trans}} = \frac{\max\{f_k(x)\} - f_k(x)}{\max\{f_k(x)\} - \min\{f_k(x)\}} \quad (11)$$

Lemma 1. For the optimal value Z^* of Eq. (1), $Z^* \in [0, 1]$ holds.

Proof. The proof of Lemma 1 is provided in Appendix. \square

Lemma 1 implies that the optimal value Z is dimensionless and lies within the interval [0,1]. Given its suitable numerical scale, further transformation in Eq. (11) is unnecessary.

Denote U_S and U_P as the weights of the objective functions of the δ -SBM and OPA models, respectively, where $U_S + U_P = 1$. Then, Eq. (10) can be transferred into weighted max-min form, as shown in

Eq. (12).

$$\begin{aligned}
 & \max_{v^-, u^+, \theta^-, \sigma^+} \min_{w, Z} \{ U_S [f_k^{trans} (\sum_{j \in J^-} (x_{ij} + \varepsilon_j^-) v_j^- + \sum_{j \in J^+} x_{ij} \theta_j^- \\
 & \quad - \sum_{j \in J^+} (y_{ij} + \varepsilon_j^+) u_j^+ \\
 & \quad - \sum_{j \in J^+} (y_{ij} - 2\varepsilon_j^+) \sigma_j^+)], U_P Z \} \\
 \text{s.t. } & \sum_{j \in J^-} x_{ij} v_j^- - \sum_{j \in J^+} y_{ij} u_j^+ \geq 0 \quad \forall i \in I \\
 & v_j^- + \theta_j^- \geq \sum_{k \in K} w_{jk} / \max_{i \in I} \{x_{ij}\} \quad \forall j \in J^- \\
 & u_j^+ + \sigma_j^+ \geq \sum_{k \in K} w_{jk} / \min_{i \in I} \{y_{ij}\} \quad \forall j \in J^+ \\
 & s_k r_{jk} (w_{jk} - w_{lk}) \geq Z \quad \forall (j, k, l) \in \mathcal{A}^1 \\
 & s_k r_{jk} (w_{jk}) \geq Z \quad \forall (j, k) \in \mathcal{A}^2 \\
 & \sum_{k \in K} \sum_{j \in J^- \cup J^+} w_{jk} = 1 \\
 & \theta_j^-, w_{jk}^l \geq 0, \sigma_j^+ \leq 0 \quad \forall j \in J^+ \cup J^-, \\
 & \quad \quad \quad k \in K
 \end{aligned} \tag{12}$$

The optimization model in max-min form can be further transformed into a linear programming problem by variable substitution. Let

$$\begin{aligned}
 \xi = \min \{ & U_S [f_k^{trans} (\sum_{j \in J^-} (x_{ij} + \varepsilon_j^-) v_j^- + \sum_{j \in J^+} x_{ij} \theta_j^- \\
 & - \sum_{j \in J^+} (y_{ij} + \varepsilon_j^+) u_j^+ - \sum_{j \in J^+} (y_{ij} - 2\varepsilon_j^+) \sigma_j^+)], U_P Z \}.
 \end{aligned} \tag{13}$$

Then, substituting Eq. (13) into Eq. (12) yields a single objective linear optimization model, as demonstrated by Proposition 1.

Proposition 1. *Given the input and output values of all DMUs, along with the variable prioritization based on a particular policy preference scenario, the δ -SBM-OPA model for CEE assessment considering policy preference is formulated as Eq. (14).*

$$\begin{aligned}
 & \max_{v^-, u^+, \theta^-, \sigma^+} \xi \\
 & \text{s.t. } U_S [f_k^{trans} (\sum_{j \in J^-} (x_{ij} + \varepsilon_j^-) v_j^- + \sum_{j \in J^+} x_{ij} \theta_j^- \\
 & \quad - \sum_{j \in J^+} (y_{ij} + \varepsilon_j^+) u_j^+ - \sum_{j \in J^+} (y_{ij} - 2\varepsilon_j^+) \sigma_j^+)] \geq \xi \\
 & U_P Z \geq \xi \\
 & \sum_{j \in J^-} x_{ij} v_j^- - \sum_{j \in J^+} y_{ij} u_j^+ \geq 0 \quad \forall i \in I \\
 & v_j^- + \theta_j^- \geq \sum_{k \in K} w_{jk} / \max_{i \in I} \{x_{ij}\} \quad \forall j \in J^- \\
 & u_j^+ + \sigma_j^+ \geq \sum_{k \in K} w_{jk} / \min_{i \in I} \{y_{ij}\} \quad \forall j \in J^+ \\
 & s_k r_{jk} (w_{jk} - w_{lk}) \geq Z \quad \forall (j, k, l) \in \mathcal{A}^1 \\
 & s_k r_{jk} (w_{jk}) \geq Z \quad \forall (j, k) \in \mathcal{A}^2 \\
 & \sum_{k \in K} \sum_{j \in J^- \cup J^+} w_{jk} = 1 \\
 & \theta_j^-, w_{jk}^l \geq 0, \sigma_j^+ \leq 0 \quad \forall j \in J^+ \cup J^-, \\
 & \quad \quad \quad k \in K
 \end{aligned} \tag{14}$$

The optimal solution set $(v_j^{*-}, \theta_j^{*-}, u_j^{*+}, \sigma_j^{*+}, w_{jk}^*)$ from Proposition 1 using the max-min method may not all be efficient. Nevertheless,

at least one element of this set is an efficient point for the multi-objective optimization model described in Eq. (10) [31,47]. After solving Eq. (14), Eqs. (4)–(8) can be employed to calculate the best technical efficiency score, lower and upper bound of efficiency, and sensitivity score of DMU l under specific policy preference. Ultimately, policymakers can cluster DMUs based on the weights of inputs and outputs, facilitating an analysis of consolidating efficiency targets within each category. This approach helps to develop the most effective developmental path for DMUs under specific policy preference.

When applying the proposed δ -SBM-OPA model, several key considerations must be addressed. First, policymakers should define the research subjects (i.e., DMUs) and their corresponding inputs and outputs. The model is applicable to DMUs of various scales, such as industries, cities, provinces, or countries. In determining DMUs and inputs and outputs, traditional DEA models impose a key requirement: the number of DMUs must meet an empirical relationship with the number of inputs and outputs, such as $I > 2 \times J^- \times J^+$ or $I > 3 \times (J^- + J^+)$. However, the proposed model, inheriting the strengths of the δ -SBM model, can handle situations where the number of DMUs is fewer than the number of inputs and outputs, meaning it does not require adherence to the traditional DEA constraints on DMU and input-output variable quantities. Furthermore, policymakers need to clearly define the policies to be analyzed, ensuring that these policies can influence DMU efficiency. Additionally, policymakers must provide a ranking of policy importance and assess the significance of inputs and outputs under different policy scenarios. Section 4 will present a case study using the ICEE analysis of Chinese provinces to demonstrate the proposed model.

4. Illustrative demonstration of industrial sector in Chinese provinces

This section demonstrates the application of the δ -SBM-OPA model through an ICEE analysis of 30 Chinese provinces in 2021, covering data collection, policy scenario analysis, and result interpretation.

4.1. Data collection

This study selects capital, labor, energy, and technology as input variables, with industry output and CO2 emissions as output variables [48]. Compared to the common inputs and outputs utilized in other ICEE studies, this study introduces technology as a novel input variable, thereby facilitating a more comprehensive ICEE analysis. As technology develops, the incorporation of the technology factor into production processes and energy consumption has the potential to significantly impact carbon emissions. The following presents the data collection and processing of inputs and outputs.

Input variable. Considering the influence of capital input characteristics on industrial sector output, this research focuses on capital stock (K), specifically using net fixed assets of large industrial enterprises as a proxy [49]. Unlike using the perpetual inventory method, this study avoids assumptions about depreciation rates, which are often arbitrarily set around figures like 9.6% or 6%. Labor (L) represents the average number of employees in industrial enterprises above the designated size [50]. About 95% of the carbon dioxide produced by human activities comes from the use of fossil fuels. Energy (E) is therefore represented by the final consumption of the eight primary fossil fuels, converted into standard coal equivalent. These fuels include hard coal, coke, crude oil, petrol, kerosene, diesel, heating oil, and natural gas. Internal expenditure on R&D by industrial enterprises above the designated size represents technology (T) [51].

Output variable. The primary business income (Y) of industrial enterprises above the designated size represents industrial output. Notably, most studies do not use gross industrial output value as a measure of industrial output value, mainly because the Chinese Industrial Economic Statistical Yearbook stopped reporting data on gross industrial

Table 1
Descriptive statistics of inputs and outputs.

| Index | Unit | Observations | Min | Max | Mean | Std.Dev |
|----------|-------------------------|--------------|----------|-------------|------------|------------|
| <i>L</i> | 10 ⁴ persons | 30 | 11.520 | 1354.170 | 264.943 | 295.491 |
| <i>K</i> | 100 million RMB | 30 | 1391.000 | 35,789.800 | 12,619.230 | 8347.020 |
| <i>T</i> | 100 million RMB | 30 | 13.849 | 2902.185 | 583.726 | 727.053 |
| <i>E</i> | 10 ⁶ tons | 30 | 0.625 | 113.427 | 32.842 | 24.624 |
| <i>Y</i> | 100 million RMB | 30 | 2676.140 | 173,649.700 | 43,804.720 | 41,142.540 |
| <i>C</i> | 10 ⁶ tons | 30 | 1.351 | 323.748 | 90.161 | 69.585 |

Table 2
Ranking of inputs and outputs under different policies.

| Policy | <i>L</i> | <i>K</i> | <i>T</i> | <i>E</i> | <i>Y</i> | <i>C</i> |
|------------------------------------|----------|----------|----------|----------|----------|----------|
| Economic priority policy (P1) | 4 | 2 | 3 | 5 | 1 | 6 |
| Environmental priority policy (P2) | 3 | 4 | 6 | 2 | 4 | 1 |
| Technological priority policy (P3) | 4 | 3 | 1 | 6 | 2 | 5 |

output value in 2012 [28]. For carbon dioxide (CO₂) emissions (*C*), given that there is no direct access to industrial CO₂ emissions by region from any statistical review or database. This study uses the method of the International Panel on Climate Change to estimate industrial CO₂ emissions across 30 provinces of China in 2021 [52]. The formula for measuring carbon dioxide emissions *CE* follows:

$$CE = \sum_{i=1}^8 CE_i = \sum_{i=1}^8 E_i \times NCV_i \times CEF_i \times COF_i \times 44/12, \quad (15)$$

where *i* is the index of fossil fuel type, and *CE_i*, *E_i*, *NCV_i*, *CEF_i*, and *COF_i* represents the carbon dioxide emissions, consumption, average lower heating value, carbon content per unit calorific value, and carbon oxidation rate of fossile fuel *i*, respectively. This study adjusts the carbon emission factors according to the National Development and Reform Commission .

This study uses data from the 2021 China Industrial Statistical Yearbook, the China Energy Statistical Yearbook, the China Provincial Statistical Yearbook, and the China Science and Technology Statistical Yearbook to analyze ICEE in 30 Chinese provinces. It should be noted that the data related to Tibet, Hong Kong, Macao, and Taiwan are not discussed in this paper, as some of the data for these regions are missing. Table 1 shows the descriptive statistics of the data.

4.2. Policy preference analysis and setting

This study introduces three policy scenarios in response to the recent policy focus of China: economic, environmental, and technology priorities. inputs and outputs within each scenario are prioritized based on policy characteristics. Certain elements are clearly prioritized under each policy preference, while others remain uncertain. Hence, correlation analysis is employed in this study to rank the importance of these elements within each policy. Specifically, this study utilizes data from the primary element under each policy across 30 provinces as a reference series. Pearson correlation tests are conducted on the data of the other elements, ranking them within each policy based on Pearson coefficients from highest to lowest. Table 2 presents the ranking results of inputs and outputs under each policy. Subsequently, government policy preferences are formed by ranking the importance of these policies. This section demonstrates the proposed model using the policy preference scenario of “P1 > P2 > P3” as an illustrative example. Moreover, all other possible policy preference scenarios will be analyzed explicitly in Section 5.

4.3. Result analysis

The result analysis includes regional differences in ICEE and cluster analysis of provinces based on input and output variable weighting frontiers as different provinces achieve optimal efficiency.



Fig. 1. Industrial Carbon Emission Efficiency Map of 30 Chinese Provinces in 2021.

4.3.1. Regional differences in industrial carbon emission efficiency

Fig. 1 depicts the ICEE across 30 provinces of China in 2021, as calculated by the proposed δ -SBM-OPA method. The results demonstrate that the mean value of ICEE across the 30 provinces is 0.623, with only 12 provinces exceeding the national average efficiency level. The provinces of Beijing, Shanghai, Guangdong, Jiangxi, and Hainan, respectively, have the highest efficiencies, all exceeding 0.990. The ICEE in Guizhou, Yunnan, Shaanxi, Ningxia, Hebei, Shanxi, Liaoning, Heilongjiang, Anhui, Shandong, Henan, Hubei, and Gansu is notably poor, with values below 0.500. The above indicates that the overall ICEE across China's 30 provinces is low and exhibits considerable variation.

This study categorizes the provinces into eight economic regions divided by the State Council of China. The mean and variance of efficiencies across the provinces involved in the eight regions and the efficiencies for each province are presented in Table A.2. The East and South Coasts exhibit the highest ICEE, with average values of 0.752 and 0.989, respectively. The following are the North Coast, Middle Yangtze, Northeast, and Northwest, with the ICEE of 0.646, 0.601, 0.597, and 0.562, respectively. However, the ICEE of the Southwest and Northwest is notably deficient, with values below 0.5000, at 0.498 and 0.487, respectively. As for the variance of regional ICEE, the East and South Coasts show quantum differences from the other regions, especially the South Coast at 0.001. In contrast, the variance in the other regions is within the interval [0.022,0.061]. As for provincial efficiencies within each region, the results show that all regions except the Southwest, Northwest, and Middle Yellow River regions have at least one province with an efficiency of 0.900 or higher. Noticeably, the three provinces on the South Coast (i.e., Fujian, Guangdong, and Hainan) all have efficiencies of 0.974 and above. This illustrates that the South Coast performs exceptionally well in terms of ICEE and has the potential to become a national benchmark for ICEE.

Subsequently, this study calculates the sensitivity of ICEE for each province within the specified uncertainty interval. Fig. 2 depicts the technical efficiency, upper and lower bounds, and sensitivities of ICEE across 30 provinces of China in 2021. The calculations show that the

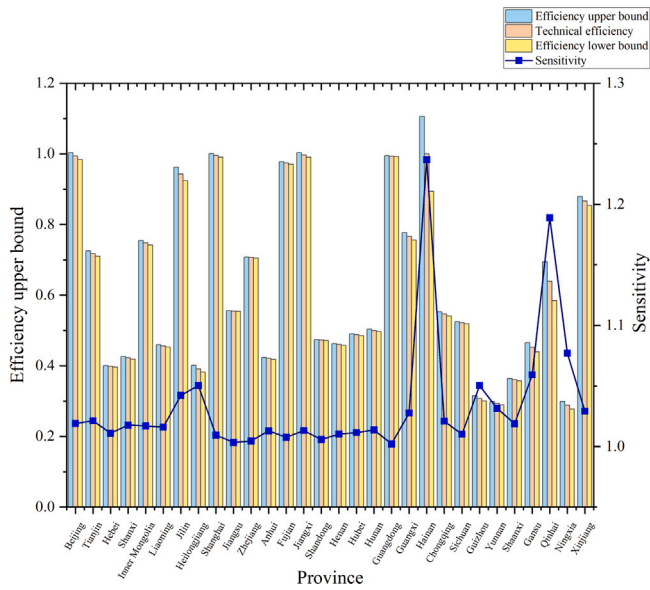


Fig. 2. Sensitivity of province carbon emission efficiencies.

mean and standard deviation of the ICEE sensitivity are 1.035 and 0.052, respectively. Notably, Hainan and Qinghai exhibit the highest sensitivity, reaching 1.237 and 1.189, respectively. These values exceed the mean plus double the standard deviation, indicating that these two provinces are particularly susceptible to data uncertainty. Even though the ICEE of Hainan has reached the efficiency frontier, its sensitivity score shows that there is still potential for further improvement in its efficiency level. In addition, the sensitivities of Ningxia, Heilongjiang, Guizhou, Gansu, and Jilin are higher than the national average. Heilongjiang, Guizhou, Gansu, and Ningxia also have low ICEE. Overall, evaluating ICEE is not the sole criterion in the context of δ -SBM-OPA. Instead, it is essential to consider the uncertainty-oriented sensitivities of each province comprehensively. Only higher ICEE accompanied by more stable results can be considered efficiency targets.

4.3.2. Cluster analysis based on variable weighting frontier

This study uses the K-means clustering method, determining the optimal cluster number based on the elbow rule. Table 3 shows the clustering results. The clustering result displays a profile coefficient of 0.707, a DBI of 0.318, and a CH of 162.582, indicating a favorable clustering effect. Among them, 30 provinces are classified into three categories with proportions of 63.3%, 26.7%, and 10.0%, respectively. The variability in inputs and outputs indicates significant differences across clustering categories at the p -value of 0.000***. This study named the three categories as technology-driven provinces (TDP), development-balanced provinces (DBP), and transition-potential provinces (TPP) based on the centroid characteristics of the weights of the inputs and outputs in each cluster, as described in Table 4.

TDP covers Tianjin, Hebei, Neimenggu, and 14 other provinces. From the weights of each input and output variable, technical inputs are the most critical factor in evaluating ICEE among these provinces, with a weight of 0.512. Capital inputs, labor inputs, and industrial output follow it. For provinces in this category, technological innovation is essential to boosting industrial productivity and reducing carbon emissions. At the same time, the focus on industrial output in these regions demonstrates their necessity to balance the need to safeguard a certain economic output level in the emission reduction process. Thus, this category is defined as a technology-driven province. Notably, among these regions, Neimenggu performs the best in ICEE and can be regarded as a benchmark of technology-driven provinces.

DBP includes eight provinces, including Beijing, Shanghai, and Jiangxi. These provinces are relatively close to each input and output ratio, with fluctuations ranging from 0.1 to 0.28. This indicates that the DBP region shows a balanced development in capital, labor, technology, and industrial output and is therefore classified as the development-balanced province. These regions emphasize the rational use of integrated resources, including labor, capital, technology, and energy, by optimizing the allocation of resources to achieve efficient operations and actively reducing carbon emissions. The ICEE values are generally high among the development-balanced provinces. Beijing, Shanghai, Jiangxi, Guangdong, and Hainan all have efficiency values over 0.99, which can be regarded as the benchmark provinces of the development-balanced provinces.

TPP includes the three provinces of Shanxi, Guangxi, and Xinjiang. These provinces have disadvantages in labor, capital, and energy consumption and lack of attention to carbon emission outputs, but are prominent in technological inputs and industrial outputs. Thus, it is clear that these provinces are now focusing on upgrading their technological inputs to promote technological innovation and improve the output quality. However, there is still a need to focus on using resources efficiently and protecting the environment in economic development, significantly reducing carbon dioxide emissions. Given this, this paper considers these provinces as transition potential provinces. Among them, the ICEE of Xinjiang is relatively excellent and can be regarded as an exemplary province of transition potential.

5. Discussion

This section examines ICEE and clustering results based on weights of inputs and outputs across various policy preference scenarios. It also presents specific recommendations for carbon emission reduction policies and strategic measures tailored to different provinces. Methodologically, this analysis serves as a sensitivity assessment of ICEE under varying policy preferences. The three policies given in Section 4.2 is permuted to generate six policy preference scenarios, outlined in Table 5. Initially, the ICEE under these scenarios is computed through the same process shown in Section 4. Subsequently, the cluster analysis on the weight characteristics of inputs and outputs is conducted to identify provinces exhibiting similar efficiency frontier features when achieving the optimal ICEE in different contexts.

Fig. 3 illustrates the ICEE under different policy preference scenarios. The results show that the ICEE across provinces under different policy preference scenarios are generally similar. The national average efficiency is highest under S5 at 0.627 and lowest under S2 at 0.621. However, as seen from Fig. 4, there are significant differences in the sensitivities of ICEE across policy preference scenarios for the 30 provinces in response to input data uncertainty. Hainan shows the most significant discrepancy at 21.5%, followed by Qinghai at 11.4%, while Guangdong shows the most minor discrepancy at 0.3%.

Table 6 shows the best policy preference options of provinces in each category and their corresponding efficiency performance. Notably, S1, S2, and S3 are not the best efficiency policy preference options for any province. In the technology-dominant policy preference scenario (i.e., S4 and S5), most provinces achieve the optimal level of ICEE, and S4 and S5 cover 23% and 67% of the total number of provinces, respectively. Under the environment-dominant policy preference scenario, Chongqing, Ningxia, and Hunan achieve optimal efficiency. Through K-means clustering analysis of indicator weights, regional groupings of provinces sharing similar efficiency frontiers is identified. Comparative analysis of optimal efficiency scores within each group pinpoint leading benchmark provinces in each category. The regional type of provinces is the same as in Section 4.3. The TDP includes 18 provinces, such as Tianjin and Chongqing, whose average optimal efficiency score is 0.543. Provinces above this average include Chongqing, Tianjin, Neimenggu, Jilin, Zhejiang, and Fujian. The TDP provinces have coverage ratios of 11%, 28%, and 61% under optimal

Table 3
K-means clustering results.

| Input and output variables | Clustering categories (mean ±standard deviation) | | | F | P |
|----------------------------|--|-----------------------|-----------------------|----------|----------|
| | Category 1 (n = 19) | Category 2 (n = 8) | Category 3 (n = 3) | | |
| <i>L</i> | 0.105 ± 0.024 | 0.121 ± 0.006 | 0.047 ± 0.005 | 15.621 | 0.000*** |
| <i>K</i> | 0.117 ± 0.017 | 0.173 ± 0.009 | 0.050 ± 0.008 | 79.475 | 0.000*** |
| <i>T</i> | 0.512 ± 0.017 | 0.183 ± 0.043 | 0.440 ± 0.004 | 442.120 | 0.000*** |
| <i>E</i> | 0.063 ± 0.032 | 0.109 ± 0.005 | 0.046 ± 0.004 | 10.126 | 0.001*** |
| <i>Y</i> | 0.162 ± 0.037 | 0.276 ± 0.017 | 0.369 ± 0.026 | 74.218 | 0.000*** |
| <i>C</i> | 0.042 ± 0.000 | 0.138 ± 0.007 | 0.048 ± 0.006 | 1845.700 | 0.000*** |

Note: ***, **, * represent 1%, 5%, and 10% significance levels, respectively.

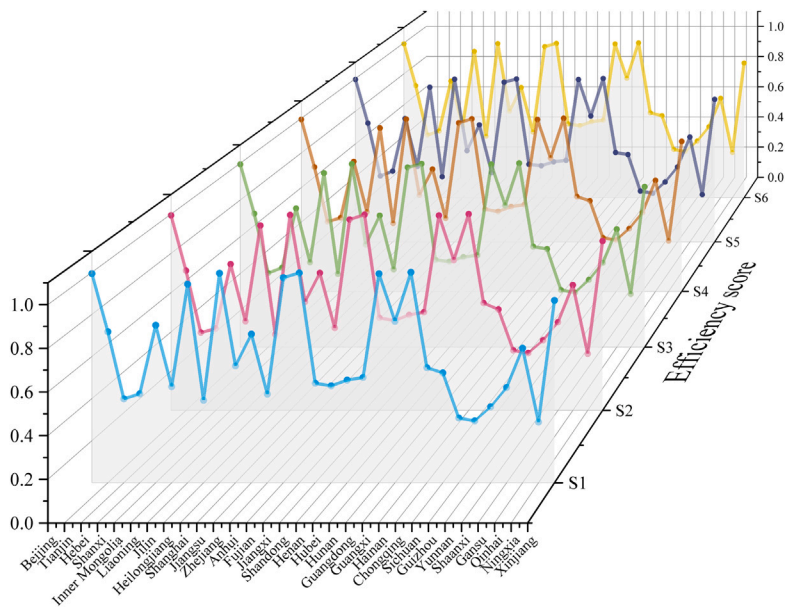


Fig. 3. Technical efficiency of provinces under different policy preference scenarios.

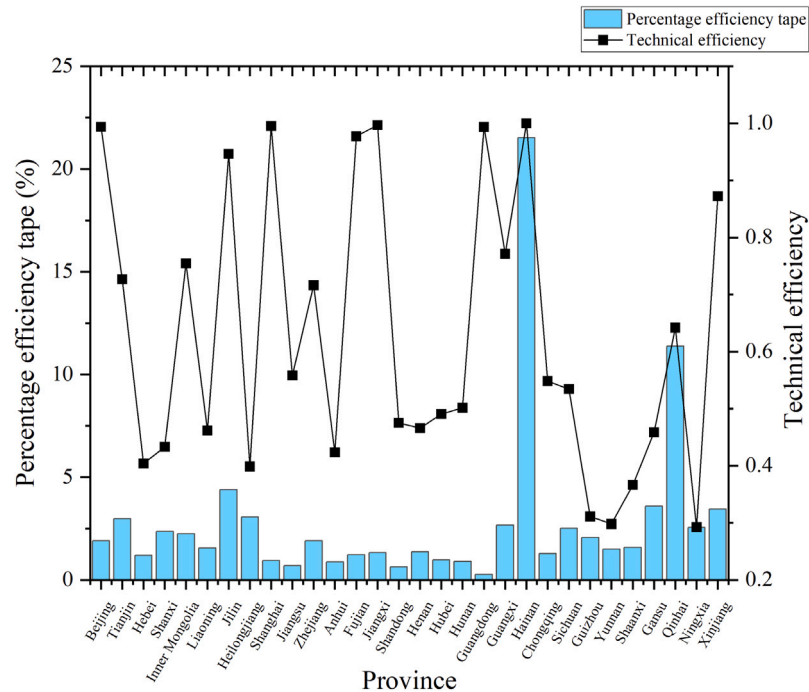


Fig. 4. Optimal technical efficiency and percentage tape across different policy preference scenarios.

Table 4
Province classification.

| Clustering category | Province |
|-------------------------------------|--|
| Technology-driven province (TDP) | Tianjin, Hebei, Neimenggu, Liaoning, Heilongjiang, Jiangsu, Zhejiang, Anhui, Shandong, Henan, Hubei, Hunan, Gansu Chongqing, Sichuan, Guizhou, Yunnan Shaanxi, Ningxia |
| Development-balanced province (BDP) | Beijing, Jilin, Shanghai, Fujian, Jiangxi, Guangdong, Hainan, Qinghai |
| Transition-potential province (TPP) | Shanxi, Guangxi, Xinjiang |

Table 5
Policy preference settings.

| Policy preference | Policy ranking |
|-------------------|----------------|
| S1 | P1 > P2 > P3 |
| S2 | P1 > P3 > P2 |
| S3 | P2 > P1 > P3 |
| S4 | P2 > P3 > P1 |
| S5 | P3 > P1 > P2 |
| S6 | P3 > P2 > P1 |

Table 6
Optimal policy scenarios and corresponding efficiency for provinces under each category.

| | S4 | S5 | S6 |
|-----|--------------------|---|---------------------------------------|
| TDP | Chongqing, Ningxia | Tianjin, Neimenggu, Jilin, Liaoning, Heilongjiang, Jiangsu, Zhejiang, Fujian, Sichuan, Guizhou, Gansu | Hebei, Anhui, Shandong, Hubei, Yunnan |
| BDP | Hunan | Beijing, Shanghai, Jiangxi, Henan, Guangdong, Hainan, Shaanxi | Qinghai |
| TPP | – | Shanxi, Xinjiang | Guangxi |

policy scenarios S4, S5, and S6, respectively. Notably, Fujian, with an optimal efficiency score of 0.977, stands out as the benchmark province in the TDP category. The BDP includes 9 provinces, such as Hunan, Beijing, and Shanghai, with an average optimal efficiency score of 0.7729. Provinces above this average include Beijing, Shanghai, Jiangxi, Guangdong, and Hainan. The BDP provinces have coverage ratios of 11%, 78%, and 11% under optimal policy scenarios S4, S5, and S6, respectively. Beijing, Shanghai, Guangdong, and Jiangxi provinces achieve the optimal ICEE values close to 1 and are recognized as benchmarks. The TPP includes three provinces: Shanxi, Xinjiang, and Guangxi, whose average optimal efficiency score is 0.692. Except for Shanxi, the ICEE of all other provinces are higher than this average. The ratio of TPP provinces under policy preference scenarios S5 and S6 is 67% and 33%, respectively. Xinjiang becomes the benchmark province in this category, with an efficiency score of 0.872.

Based on the above results and the current economic status of each province, this study proposes policy recommendations for advancing the dual-carbon strategy in three categories of provinces:

- TDP: Provinces in TDP should actively promote research and development of low-carbon technologies, using government funding to adopt efficient production technologies and innovative processes to reduce energy consumption and emissions. Enterprises should invest in energy-saving equipment and intelligent control systems to improve energy use efficiency. At the same time, it should increase investment in clean energy sources such as solar and wind and reduce its dependence on fossil fuels. It should

also introduce foreign energy-saving technologies and techniques to improve industrial energy efficiency through technological innovation. Technology-driven provinces such as Ningxia and Chongqing should focus on advanced technologies and pay special attention to the sustainable use and conservation of water, land, energy, and natural resources. Ningxia’s industrial sector has low ICEE and needs to strengthen clean energy development and technological innovation. Ningxia and Qinghai are geographically similar, and they can establish a partnership to address environmental challenges. Coastal areas such as Jiangsu, Zhejiang, and Fujian are suggested to utilize offshore wave and wind energy to accelerate the construction of a clean, low-carbon, safe, and efficient multi-energy supply system. In high-carbon regions such as Neimenggu, Gansu, Jilin, Heilongjiang, and Guizhou, local governments should promote the innovation and application of new technologies, guide enterprises to focus on the development of new and emerging technology industries and adjust the energy structure to promote the substitution of fossil energy with cleaner, renewable and non-carbon energy sources. Provinces like Hebei, Anhui, Shandong, Hubei, and Yunnan traditionally depend on abundant local energy resources. In the future, these regions must optimize their industrial structure, enhance energy efficiency, unlock emission reduction potential, and foster green regional economic growth and energy-saving practices.

- BDP: Provinces in BDP aiming for balanced development must adopt comprehensive strategies to reduce industrial CO2 emissions. Adjusting policies for energy-intensive industries will optimize resource allocation and foster regional economic complementarity. Second, upgrading industries and maximizing resource utilization will enhance CEE, exemplified by national park construction and pilot projects. Key provinces like Beijing, Shanghai, and Guangdong should leverage technological innovation to boost their economies and achieve harmonious economic and environmental coexistence. These regions should collaborate on technology and utilize innovative resources to establish an innovation-driven economic system, setting benchmarks for development. Beijing and Guangdong can lead in implementing comprehensive emission controls, while other areas should promote energy-saving technologies to transition to low-carbon industries and energy sources. Concurrently, enhancing partnerships with Guangxi and other regions in clean services will support upgrading low-carbon technologies and optimizing industrial structures across central, western, and northern regions. The technological and industrial revolution enables the central region’s transition to green and sustainable development. Henan and Shaanxi should enhance investment in industrial technology, expand clean energy supply, and effectively manage energy and water resources. Qinghai must consider the environmental impact of economic development and prevent environmental damage during resource exploitation. Hunan should speed up industrial upgrading, shifting towards technology- and capital-intensive industries while fostering the growth of low-carbon and green sectors.
- TPP: Provinces in TPP, including Shanxi, Xinjiang, and Guangxi, differ economically from the eastern region and must prioritize future economic growth. They should actively promote the dual-carbon strategy to foster a low-carbon economic model. Addressing carbon-intensive industries through technological innovation and industrial upgrading is crucial for sustainable development. Xinjiang should utilize its abundant wind and solar resources to develop renewable energy industries as a production base. Shanxi and Guangxi should transition from traditional energy to green energy chemicals through innovation, industrial optimization, and enhanced energy efficiency. Developing the green coal chemical industry, advancing clean and efficient coal power technology, and eco-friendly coal mining are vital for achieving economic

benefits and reducing carbon emissions. Coordination with regional development plans, considering differences in resource endowment and energy infrastructure, is essential for establishing a regional low-carbon spatial synergy development pattern.

6. Conclusion

This study introduces a hybrid model combining δ -SBM and OPA to analyze CEE under policy preferences and data uncertainties. The key contribution of this research is the innovative application of the hybrid δ -SBM-OPA model, which offers a more reliable and adaptable framework for assessing efficiency in DMUs across various policy scenarios. Unlike traditional DEA models that rely on predetermined weights, the proposed model adaptively determines efficiency using objective data and subjective human judgment, making it especially useful for policymakers by accommodating varying policy preferences. A significant advancement is the incorporation of policy preferences as scenario constraints, optimizing the determination of weights and enabling more customized efficiency measurements. Additionally, the proposed model accounts for data uncertainties by calculating sensitivity indicators, providing deeper insights into efficiency differences and identifying the most efficient provinces for carbon emission reduction. The practical utility of the proposed model is demonstrated by its ability to simulate the impact of different policy preferences, with K-means clustering revealing groups of provinces with similar characteristics for targeted benchmarking and policy recommendations, ultimately guiding effective carbon reduction strategies at the provincial level.

The ICEE analysis of 30 Chinese provinces illustrates the proposed δ -SBM-OPA model. This study further examines the impact of economic, environmental, and technological policy priorities on ICEE and provides policy recommendations for carbon emission reduction. The key findings from illustrative application, which provide managerial insights, indicate that:

- Regarding the policy preference of “economy > environment > technology”, the average ICEE across 30 provinces in China is 0.623, with only 12 provinces exceeding this value. The top five provinces, Beijing, Shanghai, Guangdong, Jiangxi, and Hainan, all have ICEE averages surpassing 0.990. Among the eight economic regions, the South Coast exhibits the highest mean ICEE at 0.9892 and the lowest variance at 0.001. The ICEE values for the East Coast, North Coast, Middle Yangtze River, Northeast, and Northwest range from 0.5620 to 0.7523. Conversely, the Southwest and Northwest show the lowest ICEE values, with 0.498 and 0.487, respectively. Except for the South Coast, the ICEE variances in the other regions range from 0.022 to 0.061, indicating certain variability.
- The results of the ICEE under different policy preferences show that 27 provinces have the optimal ICEE in the technology-dominant policy preference scenario. Chongqing, Ningxia, and Hunan show the optimal ICEE under the environment-dominant policy preference scenario. The provinces can be categorized into technology-driven, development-balanced, and transition-potential based on their ICEE characteristics. Technology-driven provinces have 18 provinces with an average optimal ICEE of 0.543, taking Fujian as the ICEE benchmark (0.977). Of these, 89% are optimal under a technology-dominant policy preference and 11% under the environment-dominant policy preference. In the development-balanced provinces, 9 have an average optimal ICEE of 0.773, with the ICEE benchmark nearing 1, including Beijing, Shanghai, Guangdong, and Jiangxi. This category is identical to the distribution of optimal policy preferences for technology-driven provinces, with 89% of technology-dominant and 11% of the environment-dominant. The transition-potential provinces, with an average optimal ICEE of 0.692, include three provinces where Xinjiang serves as a benchmark with 0.872,

and all provinces achieve optimal under the technology-dominant policy preference.

A limitation of the model formulation is the challenge of transforming the original SBM-OPA multi-objective optimization model into a solvable form. This study uses a standardization approach to eliminate dimensionality effects, converting the problem into a solvable single-objective linear programming model. However, the choice of standardization method may affect the efficiency outcome. Therefore, future research could explore how to convert OPA into a target-free feasible region and integrate it into δ -SBM to improve problem-solving efficiency. Another limitation of this study on the application perspective is that the conclusions drawn are based on a limited context. Specifically, this study utilizes single-year industrial carbon emission data from 30 provinces in China to demonstrate the proposed model, rather than employing multi-year panel data. Further applications in diverse scenarios and DMUs, including the use of multi-year data, are needed to validate the robustness and effectiveness of the proposed model under varying policy preferences and data uncertainty. In addition to addressing the limitations mentioned above, future empirical studies could apply the proposed model. For instance, by collecting and analyzing multi-year panel data, long-term trends in CEE could be explored. Given the regional differences in economic development, resource endowment, and industrial structure, optimal policies may differ substantially. Therefore, future research should consider these regional characteristics, offer tailored policy recommendations, and assess the differential impacts of policy implementation to provide more effective and targeted policy support for local governments.

CRedit authorship contribution statement

Shutian Cui: Writing – original draft, Visualization, Investigation, Formal analysis, Data curation, Conceptualization. **Renlong Wang:** Writing – original draft, Validation, Supervision, Software, Methodology, Conceptualization. **Xiaoyan Li:** Writing – review & editing, Investigation, Funding acquisition. **Xiuguang Bai:** Writing – review & editing, Validation, Supervision, Methodology, Funding acquisition.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix

See [Tables A.1](#) and [A.2](#).

Proof of Lemma 1. (1) Show that the lower bound of the optimal value Z^* is 0. By the derivation of OPA, the attribute that has a higher ranking $r + 1$ is dominated by the one with a lower ranking r , i.e., $A_{jk}^r \geq A_{jk}^{r+1}$, which is equivalent to $w_{jk}^r \geq w_{jk}^{r+1}$. For $\forall j \in J, k \in K$, $s_k r_{jk}(w_{jk}^r - w_{jk}^{r+1}) \geq 0$ and $s_k r_{jk}(w_{jk}^{r=n} - w_{jk}^{r+1}) \geq 0$ holds. The objective function is to maximize Z , then there exists $\min_{j,k} \{s_k r_{jk}(w_{jk}^r - w_{jk}^{r+1})\} = \max Z = Z^* \geq 0$ or $\min_{j,k} \{s_k r_{jk}(w_{jk}^{r=n} - w_{jk}^{r+1})\} = \max Z = Z^* \geq 0$ such that $\max Z = Z^* \geq 0$ holds. Thus, the lower bound of the optimal value Z^* is 0.

(2) Show that the upper bound of the optimal value Z^* is 1. Suppose that there exists ϵ_{jk} such that $s_k r_{jk}(w_{jk}^r - w_{jk}^{r+1}) = Z^* + \epsilon_{jk}$ and

Table A.1

Literature on non-parametric methods for assessing CEE.

| Reference | Method | Subject | Perspective | Policy preference | Data uncertainty |
|---------------------|------------------------------------|--|---|-------------------|------------------|
| Liu et al. (2023) | BCC | 16 cities in Anhui, China | Regular perspective | × | × |
| Ding et al. (2019) | CCR & BCC | 30 provinces in China | Regular perspective | × | × |
| Jiang et al. (2020) | Super-SBM | Logistics sector in China's 12 pilot regions | Strong transportation strategy perspective | × | × |
| Gao et al. (2021) | Super-SBM | Industrial sectors in China | Embodied carbon emission | × | × |
| Jiang et al. (2024) | super-SBM | 30 cities in Northwestern China | Water pollution and carbon neutrality | × | × |
| Fang et al. (2022) | Super-SBM | 42 thermal power plants in China | Microscopic perspective | × | × |
| Yue et al. (2023) | NDDF | 282 cities in China | Coordinated governance | × | × |
| Hu and Xu (2022) | Three-stage DEA | Export trade sector in China | Embodied carbon emission and coordinated governance | × | × |
| Meng et al. (2021) | Modified global meta-frontier NDDF | Manufacturing industry in China | Regular perspective | ✓ | × |
| Wu et al. (2017) | NDDF | 286 cities in China | Regular perspective | ✓ | × |
| Qu et al. (2022) | Robust DEA | 30 provinces in China | Regular perspective | × | ✓ |
| Guo et al. (2023) | INDEA | 30 provinces in China | Regular perspective | × | ✓ |
| Dong et al. (2022a) | Super-SBM | 32 developed countries | Regular perspective | × | × |
| Wang et al. (2022) | Super-SBM | 131 countries | Regular perspective | × | × |
| Wang et al. (2023) | Super-SBM | 139 countries | Regular perspective | × | × |
| Dong et al. (2022b) | Super-SBM | 32 developed countries | Regular perspective | × | × |

Abbreviations: DEA: Data envelopment analysis; BCC: Banker Charnes and Cooper's model CCR: Charnes, Cooper, and Rhodes's model; Super-SBM: Super slack-based measure; NDDF: Non-radial directional distance function; INDEA: Interval number DEA.

Table A.2

Regional characteristics of industrial carbon efficiency.

| Region | Province | Efficiency score | Mean value | Variance | Region | Province | Efficiency score | Mean value | Variance |
|--------------|-----------|------------------|------------|----------|-------------------------------------|--------------|------------------|------------|----------|
| North coast | Beijing | 0.994 | 0.646 | 0.054 | Northwestern | Ningxia | 0.288 | 0.562 | 0.046 |
| | Tianjin | 0.718 | | | | Gansu | 0.453 | | |
| | Hebei | 0.398 | | | | Qinghai | 0.639 | | |
| | Shandong | 0.473 | | | | Xinjiang | 0.867 | | |
| East coast | Shanghai | 0.995 | 0.752 | 0.003 | Northeast | Liaoning | 0.456 | 0.597 | 0.061 |
| | Jiangsu | 0.555 | | | | Jilin | 0.943 | | |
| | Zhejiang | 0.706 | | | | Heilongjiang | 0.391 | | |
| South coast | Fujian | 0.974 | 0.989 | 0.001 | Middle reaches of the Yellow River | Shanxi | 0.423 | 0.498 | 0.022 |
| | Guangdong | 0.993 | | | | Neimenggu | 0.749 | | |
| | Hainan | 1.000 | | | | Henan | 0.461 | | |
| | Guangxi | 0.767 | | | | Shaanxi | 0.361 | | |
| Southwestern | Chongqing | 0.547 | 0.487 | 0.031 | Middle reaches of the Yangtze River | Anhui | 0.421 | 0.601 | 0.053 |
| | Sichuan | 0.522 | | | | Jiangxi | 0.997 | | |
| | Guizhou | 0.308 | | | | Hubei | 0.488 | | |
| | Yunnan | 0.294 | | | | Hunan | 0.500 | | |

$s_k r_{jk}(w_{jk}^{r=n}) = Z^* + \epsilon_{jk}$ for $\forall j \in J, k \in K$. If $\epsilon_{jk} < 0$, $s_k r_{jk}(w_{jk}^r - w_{jk}^{r+1}) = \bar{Z} < Z^*$ or $s_k r_{jk}(w_{jk}^{r=n}) = \bar{Z} < Z^*$ holds, which contradicts the objective of maximizing the minimum Z . Thus, $\epsilon_{jk} \geq 0$ and there must be at least one $\epsilon_{jk} = 0$ such that $s_k r_{jk}(w_{jk}^r - w_{jk}^{r+1}) = Z^*$ and $s_k r_{jk}(w_{jk}^{r=n}) = Z^*$. Then, the cumulative sum of the last k constraints for each expert $j \in J$ in ascending order yields

$$w_{jk}^r = \frac{1}{s_k} \left(\sum_{h=r_{jk}}^n \frac{1}{h} \right) (Z^* + \epsilon_{jk}).$$

Substituting the normalized constraint yields

$$\sum_{k=1}^m \sum_{j=1}^n w_{jk}^r = 1 \Leftrightarrow Z^* \sum_{k=1}^m \sum_{j=1}^n \left(\frac{1}{s_k} \sum_{h=r_{jk}}^n \frac{1}{h} \right) + \sum_{k=1}^m \sum_{j=1}^n \left(\frac{\epsilon_{jk}}{s_k} \sum_{h=r_{jk}}^n \frac{1}{h} \right) = 1,$$

$$\Leftrightarrow Z^* = \underbrace{\left(1 - \sum_{k=1}^m \sum_{j=1}^n \left(\frac{\epsilon_{jk}}{s_k} \sum_{h=r_{jk}}^n \frac{1}{h} \right) \right)}_{\leq 1} \bigg/ \underbrace{\left(\sum_{k=1}^m \sum_{j=1}^n \left(\frac{1}{s_k} \sum_{h=r_{jk}}^n \frac{1}{h} \right) \right)}_{\geq 1}.$$

It follows that $Z^* \leq 1$, which implies the upper bound of the optimal value Z^* is 1. \square

Data availability

Data will be made available on request.

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